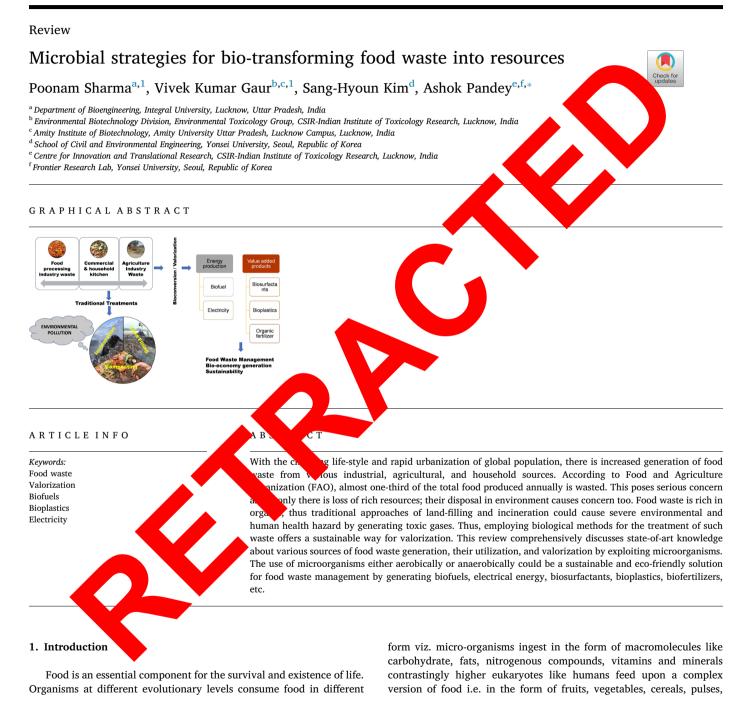


Contents lists available at ScienceDirect

BIORESOURCI

Bioresource Technology

journal homepage: www.elsevier.com/locate/biortech



* Corresponding author at: Centre for Innovation and Translational Research, CSIR-Indian Institute of Toxicology Research, Lucknow, India.

E-mail address: ashok.pandey1@iitr.res.in (A. Pandey).

¹ Both the authors contributed equally.

https://doi.org/10.1016/j.biortech.2019.122580

Received 1 October 2019; Received in revised form 6 December 2019; Accepted 6 December 2019 Available online 11 December 2019 0960-8524/ © 2019 Elsevier Ltd. All rights reserved.

Value added and bioactive compounds from diff	pounds from diff	food proc	al waste.	
Food Industry category	Waste generation tonnes)	o Ty tood pre	Value added and bioactive compounds	Reference
Fruits Processing Industry	NA	the proce.	nol, dietary fibre, grape seed oil, pomace oil, oleanolic acid, polyphenols (catechin, epicatechin,	, Galanakis 2017; Schieber 2017
		Á.	8. vid, and resveratrol), flavanols (proanthocyanidins), anthocyanins (enocyanin), procyanidins, nalates, citric acid, single cell protein.	
	5742	Apple, se processing industry	tin, acid, citric acid, aroma compounds, biogas, ethanol, butanol, and pectinases	Schieber 2017; Kiran et al., 2014
	NA	Citrus processing	Essential oil ene), phenolics, pectin, antioxidants, ethanol, organic acids, and flavanoids	Matharu et al. 2016
		industry		
	NA 13,532	Mango Banana	startor sterols, pherols, taminus, itavonols, xannones, autocyanus, and auxyresorcinols. Lig cellulo emicellulose, phenolic compounds (prodelphinidins, flavonol glycosides),	schieber 2017; Kiran et al., 2014 Schieber 2017; Kiran et al., 2014
	1829	Pineapple	Andure 1 and a second succession of the second	Schieber 2017: Kiran et al 2014
	NA	Berries	polyphe	Rohm et al., 2015
Vegetable processing industry	62,229	Potatoes		
	12.874	Tomatoes	Carotenoido vcor	Kiran et al., 2014: Schieber 2017
		Carrot	ېر بو	Schieber 2017
	5891	Onions	, phenolic c	Kiran et al., 2014; Schieber 2017
Cereal and Pulses industry	NA	Wheat	ariv	, Schieber 2017
	36 738	Rice hran	amino acids and peptides Devreius livids distary files devals antiovidant tamin F and ovyzanol)	Kiran et al 2014: Schieher 2017
	2735	Legumes	in ds. fat dds. vitamins.	Kiran et al., 2014: Parate and Talib. 2015
Beverage Industry	105	Coffee	e, starch, lipids, oteir vig	Murthy and Naidu, 2012; Kiran et al., 2014
	NIA	Ë	acid, etilation, biogas, uyes, and meta ores (centuose, nemices), ngum, pecum, gums)	Sui at al 2010
Dairy Industry	16,560	Milk	contents, polyphenolos, unaccontentol, and Biodiesel, ethanol, whey protein, Lactose, cast, and mine	Kiran et al., 2014; Parashar et al., 2016;
Meat and Seafood Drocessing	1344 F	Meat and noultry	Bertilizer animal feed blood meat and hone meal feed and production	Lappa et al., 2019 Viran et al. 2014: Ninc et al. 2018: Vaakoh
Industry	0			et al., 2019
	NA	Fishes	Chitosan, and glycosaminoglycans,	Prameela et al., 2017
	6000 to 8000	Shrimps	Nutraceuticals (Astaxanthin), chitin and chitinase	Kiran et al., 2014, Yan and chen, 2015;
		-		Prameela et al., 2017; Kumar et al., 2018
		Lobster Lobster	Chitin, calcium carbonate and protein Chitin	Yan and chen, 2015 Yan and chen. 2015
Edible oil Industry	21,462	Oil	Biosurfactants like rhamnolipids and glycolipids, biodiesel, tocophero. rols, squalene, ap	Henkel et al., 2012, Kiran et al., 2014
		Olives	Phenolic compounds, polyphenols, carotenoids, phytosterols, squalene, and ary fib	n et al., 2014; Schieber 2017

meat, and dairy products. A major concern arise when this indispensable commodity i.e. food is misused and mismanaged at any stage of the food life cycle leading to serious social, economical and environmental consequences. Table 1.

The leftover or precooked food which generates biodegradable organic waste is termed as food waste (FW). As per the definition given by The Food and Agriculture Organization (FAO), FW is "food losses of quality and quantity through the process of the supply chain taking place at production, post-harvest, and processing stages". Specifically FW corresponds to the loss of food at the end of food life cycle (Tsang et al., 2019). Generation of FW also leads to considerable loss of other resources like water, land, labour and energy. It was estimated by FAO that annually 1.3 billion tonnes of wasted food is generated globally. This wasted food is one-third of the total food produced globally, whose production corresponds from 28% of agricultural area utilizing 1.4 billion hectare of the world's fertile land (Karthikeyan et al., 2018; Paritosh et al., 2017). It is projected that economic and population growth will lead to increased FW generation in next 25 years in Asian countries. The urban FW is expected to rise 138 million tonnes by 2025 as compared to in 2005 (Paritosh et al., 2017). Out of the total waste generated globally, Asia contributes to highest FW 278 million tonnes, whereas Vietnam produced approximately 11.55 million tonnes of FW (Kiran et al., 2014). This FW includes fresh fruits, vegetables, dairy products, bakery products, and meat from diverse sources, including discharges from households, hospitality sector, food processing industries, commercial kitchens, and agriculture waste (Karthikeyan et al., 2018; Kiran et al., 2014).

The disease control centers prevented the use of FW as animal feed, thus preferentially the disposal of FW was done through fermentation, composting, and landfilling (Tsang et al., 2019). Conventionally, FW is an element of municipal solid waste which is dumped or incine The high moisture content of FW leads to the generation of dioxi incineration, whereas dumping in open area causes environmental health issues. FW is estimated to generate 3.3 billion tong of CO^2 r year, thereby contributing in the emission of gase (Paritosh et al., 2017). To increase environment astain ity and overcome socio-economic concerns, valorization f FW duction of value added products is an ideal ppr esearches on tion of researchers worldwide. This is als dent as valorization of FW has increased ≥ 9 uring the la ade from lerstand the 2009 to 2018. Therefore, this review rces and nature of FW that can be efficiently converted value added products such as biodiesel, ethanol, bi drogen, meth. butanol, biosurfactilizers and electri tants, bio-plastics, organi ower generation ded/dumped FW. e of di highlighting the signific

2. Sources of food waste ation

2.1. Food pr sing in stries

2.1.1. Cere

e, significant proportion of the human diet is filled On a globa by cereal grains ined from seeds of Gramineae family such as wheat, rice, barley, ze, sorghum, millet, oat, and rye. According to the FAO, globally total crop production during 2016 reached 2577.85 million tons. In contrast, the production of coarse grains (cereal grains other than wheat and rice used primarily for animal feed or brewing) was 1330.02 million tons (FAO-AMIS, 2017). The production ranking in the year 2014 was corn 1253.6, rice paddy 949.7, wheat 854.9, barley 146.3, oat 23.2, and rye 15.8 tonnes (Papageorgiou and Skendi, 2018). The cereal and pulses processing industry produce a large quantity of by-products like bran and germ, during processing (Anal, 2017). India is the worlds' largest producer of pulses producing a considerable amount of husk as a by-product during processing (Parate and Talib, 2015). Husk is recycled in many ways to produce high-end products. There are many crops which generate husk as a by-product of

processing. Apart from utilization as animal feed the straw, husk and dried leaves of crops like wheat, corn, rice, and barley is utilized in traditional way for making thatching roofs, baskets, broom, hand fans, handbags and in preparation of decorative items. It is used as cleaning and polishing agent in metal and machine industries. Rice husk can be used as pet feed fiber, fertilizer and substrate for vermicomposting technique, and in the production of construction material like light weight bricks (Kumar et al., 2013). Husk obtained after cocoa pod processing was utilized for pectin extraction and production of vermicompost, oyster mushrooms, livestock feed, and other value-added products (Dede and Ozdemir, 2018). Furthermore, coconut husk has multiple household applications like rope, by mat, tiles, fishing net, and mattresses. It is also employed ction of second generation bio-ethanol (Bolivar-Teller al., 20 These by-products are rich in nutrients, generally ts of dietary ers, proteins, lipids, fatty acids, vitamins, mine olic com inds but still ls and they arrive finally as animal fuel, and refin substrate. For production of refined flour an and rtion of grain is term. removed as they adverse ect th cessing properties (Verni et al., 2019). The by-produce of ba ling pro serves as a rich source of bioactive compa s like p es, ins e dietary fiber, phenolics than in whole barley grain and it contains imes more di, 2018). e husk is used in fermentation (Papageorgi process to adjust mol maintain the porosity of fermentative maring distillation (Tan et al., 2014). terial f ous exchai

Fruits and vegetables

2

bles are energy rich food items with high moisture uits and ve having ri utritive profile consisting of soluble carbohydrates co ruct vitamins, minerals, fibers, polyphenols and other (glue unds (Schieber, 2017). Fruit and vegetable wastes were bioactive ated during different steps of food supply chain starting from farm cluding production, processing, packaging, handling, storage nd transportation (Ji et al., 2017). Fruits and vegetables are classified to waste category only when a consumer disqualifies it from degree of acceptance. This may arise due to several factors such as discoloration, wounding or chilling, biochemical reactions (enzymes, antioxidants, phenolic compounds and oxygen), thermal treatment, microbial attack (rotting, softening and surface growth), and degree of ripening. India is the second largest producer of vegetables and fruits in the world sharing 10% and 14% of global production, respectively. It leads to economic loss worth of US \$483.9 million per year due to wastage of about 50 million tons, accounting for 30-40% of total production in India (Panda et al., 2016). Central de Abasto, the second largest fruit and vegetable market in the world, at Mexico City, produced 895 tonnes of waste/day. China produces approximately 1.3 million tonnes of this waste per day (Ji et al., 2017). According to a report of FAO in 2014, UK alone produced 5.5 million tonnes of potatoes; around 3% to 13% of harvested crop never reaches to customer and is wasted due to "grading losses" in supermarkets. This generated waste is processed by composting, land-filling, incinerating or used as animal feed. These disposal methods give rise to serious environmental concern such as toxic and greenhouse gas emission, microbial proliferation due to high content of moisture and landfill leachate (Ji et al., 2017; Dessie et al., 2018).

Reduction of fruits and vegetable waste is warranted to alleviate the increasing demand for food production and improving the overall efficiency of food supply chain (Matharu et al., 2016). Growing public concerns about hunger, fruits and vegetable losses, food security reasons, conserving the environment from pollution, socio-economic factors have accelerated research into FW domain towards finding better ways of using this natural and renewable resource. The starch, cellulose and/or hemicelluloses of fruit and vegetable waste is hydrolyzed to soluble sugars and further fermented to produce ethanol and hydrogen (Díaz et al., 2017). Microbial processing of fruits and vegetable waste has opened new horizons for value addition to rejected fruits and

vegetables. Several high-end commodities was reported to be produced by utilizing FW such as fermented beverages (fenny, vineger), singlecell proteins (*Saccharomyces* sp., *Candida utilis, Endomycopsis fibuligera* and *Pichia burtonii*), single-cell oils, polysaccharides, dietary fibre, polyphenols, bio-pigments (carotenoids), fragrances, flavours (vanillin), essential oils, biopesticides, plant growth regulators, enzymes (cellulase, amylase, protease, phytase, etc), biohydrogen, bioethanol and biogas (Panda et al., 2016; Schieber 2017; Sabu et al., 2002; Bogar et al., 2003a; Pandey and Soccol, 2000; Benjamin and Pandey, 1997). Acidogenic fermentation of fruit and vegetable wastes produces lactic acid (Wu et al., 2015) whereas in solid state fermentation they are hydrolysed using crude enzyme mixtures to produce succinic acid (Dessie et al., 2018).

2.1.3. Dairy

Around 29 million tonnes of dairy products are wasted in Europe every year. This dairy waste is derived from the processing industry, spoilage of the dairy products due to microbial attack, and inappropriate handling (Mahboubi et al., 2017). Dairy waste consists of complex organic constituents like fat, protein, sugar, traces of food additives, and detergents that are used for maintaining proper hygiene and clean in place (CIP). Dairy products are the most perishable commodities due to their rich composition and absorbability. Fungal contamination in milk produces visible or non-visible defects, such as offodor and flavor development.

India is reported as the largest milk producing nation and simultaneously produces 1 to 3 times of effluent for every volume of processed milk, thus generating 3.739-11.217 million m³ of waste per year. While manufacturing of cheese, a considerable amount of whey is produced as a by-product of processing which generated 9 kg of whey per kg of cheese produced (Parashar et al., 2016). Another constituent of produce i.e. raw milk, contaminates groundwater due to the presen ammonia, nitrogen, and nitrate that is converted to nitrite ther causing methemoglobinemia. During processing of raw milk arou 2.5-3.0 L of wastewater is generated per litre of prog (Sing et al., 2014), carrying about 14-830 mg/l of total ogen icentracount of tion (Kushwaha et al., 2011). Wastewater hold nificar nutrients, like carbohydrates, lipids, protein Mil ch as nucleic and gives rise to organic and inorganic for nitro_b 3, NH+ acids, urea, proteins, and NO- 2 spectively (Kushwaha et al., 2011). It was for evated conc ation of NO3 > 40 mg/L in groundwater is a cau methemoglobinemia (Fewtrell, 2004).

This raises serious env mental concerns and demands the emred m ployment of microbial d of waste conversion including reactor tickling filter, anaerobic activated sludge, seque ons (Di al., 2014). Dairy waste is sludge blanket, and aerated rich in organic cilita he wth of microorganisms hence ets can be obtained by utilizing of va a large nup added p tose and protein (Lappa et al., 2019). It is a waste dairy ind duction using Saccharomyces cerevisiae suitable sub alysis of fermentable sugar (Parashar et al., 2016). via enzymatic Filamentous fung luce a variety of enzymes capable of hydrolyzing complex carbohydra to simple sugar hence aids in high quality biomass production that can be used as animal / fish feed and even for human consumption as single cell protein with a GRAS status (Mahboubi et al., 2017).

2.1.4. Edible oil

Edible oil industry generate waste during various steps of refining process like degumming, neutralization, bleaching, deodorization, oxidative or hydrolytic rancidity. This hydrolytic rancidity is caused due to the oxidation of lipids, aging, moisture, presence of oxygen and effluent coming out of industry laden with lots of fatty acids, carbohydrates and protein (Okino-Delgado et al., 2017). Waste cooking oil is an oil-based substance that resulted from multiple deep fat frying process which makes fat unsuitable for human consumption due to the formation of polar compounds like free short-chain fatty acids, monoand di-glycerides, aldehydes, ketones, polymers, cyclic and aromatic compounds. It was reported that edible oil industry annually produces 350.9 million tonnes of de-oiled cake and oil meal as a by-product, which is a concentrated source of protein. After pretreatment, this waste is further utilized for preparing human nutrition products, animal feed and fertilizer (Chang et al., 2018).

Traditionally, effluents from oil processing industry were released directly into the soil and groundwater which leads to oily film formation on aquatic surface causing a serious threat to survival of aquatic animals, blockage of sewage and drains due to ulsification of organic matter, oil methanization worsening greep Okino-Delgado et al., 2017). Advance methods emp microbia ls for biodegradation of organic matter from efficient thereby pr cing various high-end products such as bio-ba nic bios ctants. Pseud zw domonas aeruginosa synthesiz iosurfact uch rhamnolipids and glycolipids, biodiesel action ing lip aid hydrocarbon Edible oil industry biofuels (Henkel et al., · Che al., 2018, waste was reported ource f ealth constructive probe squal ducts such as toc rols, stu which were used in difa for SCP production (Diwan ferent industrie w material, et al., 2018) tical formul ns and cosmetics in the form of soap stalk (Sherazi an hesar, 2016).

poultry and egg 2.1 eat poultry and egg processing industry is a huge segment of food system. Eu an Union annually accounts for approximately 11 cl tonnes o oduction. About 3.5 billion pounds of beef was mi .. c a in 2006, contributing \$26 billion to its economy. prod vast quantity of animal by-products, slaughterhouse Consequ and wastewater is generated (Ning et al., 2018). This comprises cattle, 47% from sheep and lambs, 44% from pigs and 37% rom chicken 37%, which is inedible in nature, thereby generating a huge mass of waste by slaughterhouses in environment (Adhikari et al., 2018).

Major wastes material generated in industry include feathers, hairs, skin, horn, hooves, soft meat, deboning residues, bones, etc. In addition to this, the slaughter house wastewater consists of blood residue, protein, animal fat (lard and tallow), detergent residues and high organic matter (carbon, nitrogen and phosphorous). Rendering of perishable animal waste extends a possible alternative way to eradicate the environmental issue along with revenue generation. Rendering industries produces meat and bone meal, hydrolyzed feather meal, blood meal, fish meal, animal fats (lard and tallow) (Yaakob et al., 2019). Lactic acid fermentation of slaughterhouse waste is a promising approach to utilize slaughterhouse waste for the production of lactic acid bacteria that can be used as a probiotic product (Ashayerizadeh et al., 2017). Due to high nutritional composition slaughterhouse waste can be utilized for various value-added product generation like biomass (Scenedesmus sp.) as fish feed (Yaakob et al., 2019), a clean energy substitute i.e. methane produced by anaerobic digestion of wastewater (Ning et al., 2018), biogas production from poultry litter, blood in food applications (blood sausage, blood cake, blood pudding, blood curd) and non-food application (fertilizer, binder) (Adhikari et al., 2018). Biodiesel has been produced from chicken manure biochar by pseudo catalytic transesterification reaction, from pork fatty waste by fermentation employing Staphylococcus xylosus and from eggshells by transesterification of triglycerides with methanol using homogeneous catalysts (Marques et al., 2016). It has multivariate applications in the pharmaceutical, cosmetic product development industry.

2.1.6. Sea foods and aquatic life

Marine ecosystem supplies around 20% of the world's food to humans, therefore plays a vital role in feeding a significant population of the world. It is estimated that around 183 million tonnes of seafood

would be needed by 2030 for feeding. Generally, 50-70% of seafood raw material is wasted every year (Kumar et al., 2018). Globally producing about 6-8 million tonnes of waste in the form of crabs, shrimp and lobster shells, where Southeast Asia contributes 1.5 million tonnes in totality. In seafood's the utilizable mass of marine animal is less, crabs yield 40% meat of whole mass, in tuna fishes only 75% of fillets is available. This generates a huge amount of waste, including inedible parts like shrimp shells, crab shell, prawn waste, fish scales and endoskeleton shell of crustacean. Shells and scales of aquatic animals such as shrimps, lobsters and fishes harbor useful chemicals such as protein. chitin and calcium carbonate. Dried shrimps and crabs contain about 50% of chitin on dried basis, generating a revenue of around \$100–120 per tonne (Yan and Chen, 2015), and is used as animal feed supplement. bait or fertilizer. Seafood waste is chitinaceous in nature, harbouring pathogenic microbes, carcinogens, aflatoxins and other health risks it can cause due to bioaccumulation of these contaminants.

The conventional method for seafood waste disposal includes ocean dumping, incineration, landfilling, and wastewater discharge of seafood processing industry. It contains high organic material load that potentially causes euthrophication and oxygen depletion of receiving water bodies (Yan and Chen, 2015). Chitin, the major component of seafood waste is insoluble in water and inert to most of chemical agents hence leads to environmental pollution and bio-fouling, thus stipulating biological processing of this waste material as a step toward environment protection and a sustainable way of billion-dollar revenue generation as bio-economy (Yan and Chen, 2015; Kumar et al., 2018).

Indian shrimp processing industries produces more than 0.15 million tonnes of shrimp waste annually. Waste of shrimp processing industries was used to produce nutraceutical compound such as astax-(3,3'-dihydroxy-ß-carotene-4,4'-Dione), anthin а xanthophyll carotenoid present in crustacean waste that was extracted th oxidative transformations of ingested β -carotene or zeaxanthin by microalgae (Prameela et al., 2017). Chitosan is an important bid lymer extracted from exoskeletons of crustaceans waste. It was report to exhibit antibacterial activity against Enterococcus herichi ein et coli, Staphylococcus aureus, and Candida albicans (2013) Glycosaminoglycans (GAGs) extracted from ma anima is of better quality to that of terrestrial organism rce for func-Waste of Seafood processing industries otentia tional and bioactive compounds prod through en mediated hydrolysis.

2.2. Commercial and househol

nomic elopment and uncontrolled po-Urbanization, rapid pulation growth has incl nsumption of food leading to many of kitche folds increase in the aste annually (Zhao et al., genera 2017a). In 201 fuced 200,000 kg of domestic eas o Chir produ nore than 30 million tonnes of waste. Pres every has been estimated that annually 2.5 billion kitchen w tonnes of F opean Union out of food supply chain (Li et al., 201)

Kitchen wast () is a kind of anthropogenic organic waste, usually discharged fi public catering rooms, restaurants, households, canteens of school and factory etc. (Li et al., 2017; Zhao et al., 2017b; Liu et al., 2019). The waste generated by restaurants, commercial and institutional kitchens is different from municipal solid waste and is referred to as kitchen waste (KW). KW includes fruits, vegetables, cooked food wastes, meat, used fat, oil and grease. On wet basis KW broadly comprises of fruits 38.2%, vegetables 41.5%, staple food 7.6%, egg shell bones 7.2%, shells and pits 2.5%, and meat 2.3% (Zhao et al., 2017a). Chemically it comprises of carbohydrate polymers (cellulose, hemicellulose, pectin and starch), protein, lignin, fats, organic acids, inorganic salts and others (Chen et al., 2017; Zhao et al., 2017b). KW is generated during various food processing operations such as handling, processing, production, storage, transportation and consumption.

Commercial and household Kitchen wastewater is generated during various activities like washing and rinsing foodstuffs, cooking, cleaning dishes, cooking utensils, and general housekeeping.

Traditional methods of decomposition of kitchen waste such as land-filling, composting, incineration and direct or indirect discharge of wastewater into the environmental system is detrimental to the ecosystem and human health (Chen et al., 2019). Kitchen waste is biodegradable biomass with higher moisture content and a pool of nutrients facilitating the growth of pathogenic microorganisms causing rotting and breeding of flies and mosquitoes in shorter span of time. It emits toxic substances, greenhouse gases, a huge amount of leachate in water bodies and foul smell of ammonia and hydrothi Ammonia produced during decomposition of kitchen waste has ent order which can cause serious irritation in the respi s of eyes and v tract. skin, while hydrothion is highly toxic umans (Zha al., 2017a). The uncontrolled generation and uld produce prop nagemen lethal and life-threatening co ment. It was dences o en reported that kitchen waste d be 1 rd as a e for generation of numerous high value ⁴ucts biosurfact. It production using Pseudomonas aeruging 1., 2018 butanol production by 2017 enzymatic hydroly Chen actic acid production by ilus, volatile Lactobacillus ar as and hydrogen from mixed), cellulose culture (Liu duction via Aspergillus niger in solid state fermentation hanol production by a process of successive liquefa pre-saccha tion, and simultaneous saccharification et al., 2017). It can be used to harness and ation (Nishimu y liberating products like biogas, bioelectricity, biodiesel, etc. er composites opolymers and edible film like materials can be Ν ized using as substrate. FW can be transformed to huge array sy roducts including antioxidants, pigment, nudda of ary fiber, organic acids, high fructose syrup, single-cell traceutio in, vermicompost, biofertilizer, xanthan gum, and wax esters, sisly reducing environmental burden (Liu et al., 2019).

2.3. Agricultural waste

Waste framework directive 2008/98/EC laws defines waste as "any substance or object which the holder discards or intends or is required to discard". Agriculture waste is an organic substance involving straw, bagasse, molasses, spent grains, husk (rice, maize, and wheat), shell (walnut, coconut, and groundnut), skin (banana, avocado), crop stalks (cotton), plant waste, livestock and poultry manure (Dai et al., 2018). In 2013, FAO reported that about 250 million tonnes of non-edible plant waste from different crop processing is generated as agro-industrial waste (Heredia-Guerrero et al., 2017). Being world's largest grain producer, China produced 1.75×10^9 tons of agriculture waste, of which 9.93 \times 10⁸ tonnes was contributed by crop straw, 4.52 \times 10⁸ tonnes from livestock and poultry manure, and 3.03×10^8 tonnes from forest residues in 2013 (Dai et al., 2018). Asia alone generates 4.4 billion tonnes of solid wastes annually. Agro-industrial waste generation from different sources in India is more than 350 million tonnes per year (Madurwar et al., 2013).

Traditionally, agricultural waste is either incinerated/burnt off or allowed to rot in fields, thereby causing serious air pollution and is also exacerbating soil pollution, water pollution and food contamination. These traditional approaches liberate toxic gases (such as N₂O, SO₂, CH₄,), smoke, greenhouse gases into thfe air, and other carcinogenic chemicals such as dioxins, furans and polycyclic aromatic hydrocarbons, which are detrimental to the environment as well as for human health. Exposure to these chemicals causes severe development damage in fetuses, infants, children, and adults (Cheng and Hu, 2010). Agricultural waste is biodegradable and organic, possessing a repository of nutrients such as polysaccharides (starches, cellulose, hemicellulose), proteins, lignins, fibers, minerals, vitamins, and others (Madurwar et al., 2013). Agriculture waste is porous and loose in constitution containing carboxyl, hydroxyl, and other reactive groups, so agricultural biomass can be used for wastewater remediation as an adsorption material, hence "reducing waste by waste" (Dai et al., 2018). Apart this, the chemical composition of agriculture waste proves it as a versatile candidate bearing potential to synthesize a number of products such as bioplastic from cuticles present on the outer layer of plants parts such as leaves, stem, flowers (Heredia-Guerrero et al., 2017).

Agriculture waste can be a cheap and natural substitute for production of multiple high valued products. Microbes can readily feed upon agriculture waste to generate an array of high end products such as pigments, phytochemicals, antibiotics, different enzymes (endoglucanase, β -glucosidase, amylase, glucoamylase) by utilizing peels, seeds, oil cakes, field residues, and bran (Bogar et al., 2003b; Selvakumar et al., 1998; Singh et al., 2006). Xanthan is a type of exopolysaccharide that acts as food additive in food industry and is reported to be produced by action of *Xanthomonas* species on agro waste as substrate (Sadh et al., 2018).

Food additives play a significant role in enhancing the technological properties of food. Mushroom is ecology and economy balancing crop derived of lignocellulosic agro waste (wheat paddy, rice paddy, banana leaves, cotton stalks) by mushroom fungi *Lentinula edodes* and *Pleurotus* sp. (Philippoussis, 2009). Agro-industrial waste is reported to be a good carrier for enzyme immobilization and solid state fermentation (Sadh et al., 2018).

Agricultural material can be utilized to produce suitable substitutes of construction material with improved qualities like light weight, biodegradable and eco-friendly material. It can be used to develop several construction items like fiber-board made from cotton stalk with no chemical additives, thermal insulating walls and roofs, waste-create bricks made from recycled paper mills waste and cotton waste. Cement replacer is manufactured from sugar industries waste (bagasse as oil palm shell as a coarse aggregate for structural concrete product (Madurwar et al., 2013).

3. Food waste utilization for energy production

The current strategies of landfilling or interacting and do not eliminate the environmental and economic tress and the gaining much interest. FW can be converted a different tims of energy molecules or biofuels, viz. biogas, here en, ethanologies biodiesel, butanol and methane (Table. 2).

3.1. Production of biofuels

3.1.1. Biodiesel

Increasing environm ion, fuel demand, and depletion of e requ fossil fuels have ent of alternative fuels. tensifi Biodiesel emer hat can be acquired from oils altern reen whose ction is expensive because of the and fats. It eedsto in order to reduce the cost, possibility to high cost utilize FW have been investigated (Karmee et al., converted to biodiesel by direct transesterification 2015). FW cal employing acid/a e catalyst or of microbial oils that are produced by oleaginous micro nisms (Kiran et al., 2014). Microbial oils have similar fatty acid composition as plant oil thus microbial oils can be

used as raw material for production of biodiesel. FW hydrolyzate can be used as nutrient source and culture medium for the cultivation of microalgae for biodiesel production. *Aspergillus oryzae* and *Aspergillus awamori* was used to prepare FW hydrolyzate, which served as growth medium for cultivation of *Chlorella pyrenoidosa* and *Schizochytrium mangrovei* yielding 10–20 g biomass. The fatty acids produced by *C. pyrenoidosa* and *S. mangrovei* were suitable for the production of biodiesel (Pleissner et al., 2013). Waste cooking olive oil was reported to serve as a substrate by *Penicillium expansum* and five different strains of *Aspergillus* sp. for the production of biomass rich in lipid. Among these, *Aspergillus* sp. ATHUM 3482 yielded highest amount of lipid 0.64 g/g dry cell weight (Papanikolaou et al., 2011). Biodiesel production from vegetable oils, animal fats and butter was reported to globally yield 24.5 GJ energy per year (Kiran et al., 2014). Cubas et al. (2016) reported biodiesel production employing corona discharge plasma reactor technology (CDPT). This technique offers several advantages such as increased esterification, easy biodiesel separation, and no co-product generation. CDPT has been efficiently used for biodiesel production from waste frying oil in the absence of chemical catalyst.

3.1.2. Ethanol

Wide industrial application of ethanol has increased its demand globally. It is prominently used as a feedstock above a feedstock above a set of the set ethylene (Kiran et al., 2014) and an ess aterial for production of polyethylene. Traditionally, lose and hy crops, e.g. sugar cane, rice, and potato were u or producti f bio based ethanol (Thomsen et al., 2003). available to mme nzymes convert starch to glucose wh is furthe o ethanol by nent Saccharomyces cerevisiae, w as celly se hyd s difficult (Kiran et al., 2014). Thus FW low g osic conte a serves as a better alternative for the pro ucti thanol improve the purity and yield of ethanol, s autod prior ermentation, as harsh or thermal treatm ay cause pa adation of nutritional comkai and Ez 2006). Dried FW has reduced ponents and surface area which i in decreased reaction efficiency between substr enzyme, t resh and wet FW was reported to be an effi roduction. Furthermore as cellulose or arce for ethano. cannot be directly converted to ethanol by yeast thus the const on efficienc dependent on carbohydrate saccharification (Kiran v s a mixture of enzymes namely glucoamylase, α et 2014). For amy and pullulanase etc. will be added for high molestrates. Hong and Yoon (2011) reported the production cular we g ethanol and 60 g reducing sugar in 48 h of fermentation by 00 g of FW. Kitchen garbage and non-diluted FW was reorted to yield 17.7 g/L/h and 0.49 g ethanol per g sugar by simultaneous saccharification and fermentation process (Ma et al., 2009).

Several countries including Japan, Finland and Spain had developed

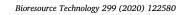
pilot scale plants to convert their FW to bioethanol (Kiran et al., 2014).

3.1.3. Biohydrogen

Hydrogen is a potential renewable energy source with high energy yield of 142 MJ/kg (Jarunglumlert et al., 2018). In nature, hydrogen is not readily available but can be produced from primary energy source. Hydrogen being carbon free with the highest energy output and yielding water on combustion is globally used as a substitute to fossil fuels (Nikolaidis and Poullikkas, 2017). Biohydrogen is hydrogen gas produced employing biological processes. The production of biohydrogen through fermentation using FW requires less energy and simultaneously recycles the waste generated, thus it has become the most favored method. Carbohydrate-rich FW is preferred for the production of biohydrogen gas as it has 20 time higher potential than protein or fatbased FW. Hydrogen production from FW had been reported employing several fermentation processes such as continuous, semi-continuous, batch, single or multiple stages (Kiran et al., 2014). Upflow anaerobic sludge blanket (UASB) and anaerobic sequencing batch (ASBR) reactors have been used with high hydrogen producing rates because of their high biomass concentration in the reactor (Karthikeyan et al., 2018). For these processes, solid retention time (SRT), hydraulic retention time (HRT), and organic loading rate (OLR) affect the production of hydrogen. An optimized SRT with low HRT enhanced the technical feasibility and productivity of biohydrogen production. However, the role of a higher OLR is yet debatable as it affects the process in both ways, suggesting an optimal OLR for maximum biohydrogen production (Kiran et al., 2014).

As in anaerobic digestion, the process of dark fermentation is set at low HRT with high acidic condition to inhibit methanogenic activity. The dark fermentation process for the production of biohydrogen from

P. Sharma, et al.	



Siologic	Table 2 Biological process for valorization of food waste. c No Substrate	ie. Bioloriieal Drei, secoristad	thes Asconistical	Devoluct Ohtsingd	Roferent.coc
o. No.			A des Associated	Product Obtained	kererences
1.	Noodle Waste	Saccharification and ferment	lar re	Bioethanol and Biodiesel	Yang et al., 2014
ci c	Waste cooking olive oil	Fermentation	Penio rasum	Biodiesel	Papanikolaou et al., 2011
'n	maxeu roou waste (rouato peet waste, Banana Peel, Household waste)		mylase;5	DIOCULATION	Valifice, 2010
4	Agricultural residue waste	Enzymatic hydrolysis and Flash Pvrolvsis	div ¹ rase, p se and lipase	Hydrochars and bio-oil	Karmee, 2016
ιċ	Cooking oil waste	Enzyme immobilization	Lipase	Biodiesel	Heater et al., 2019
6.	Spent coffee grounds	Extraction and transesterification	Catal (OH)	Biodiesel	Kondamudi et al., 2008
	Cane molasses and starch rich food waste	Fermentation	Clost n acetobutylic um beijerinckii P260	Biobutanol	Girotto et al., 2015; Huang et al., 2015: Uior et al.,
œ.	Mixed food Waste	Microbial electrolysis cell (MEC) and Anaerobic Digestion	Exoelecte	Methane	Park et al., 2018
10.	Carbohydrate rich food waste	Fermentation,	Up-flow anaerobic sh ket (UASR) and anaerobic sequencing batch (ASB) ctors	Biohydrogen	Kiran et al., 2014, Karthikeyan et al., 2018
11.	Industrial wastewater domestic wastewater and excess sludge	Microbial Fuel Cell		Electricity	Li et al., 2016; Cercado-Quezada et al., 2010
12.	Kitchen waste oil	Fermentation	Pseudomonas aeruginosa	Biosurfactant	Chen et al., 2018
13.	Soy molasses	Fermentation	Pseudomonas aeruginosa ATCC 1 45	Glycolipid biosurfactant	Rodrigues et al., 2017
14.	Dairy effluent (cheese whey)	Fermentation	Pseudomonas aeruginosa SR17	Biosurfactant	Patowary et al., 2016
15.	olive oil mill waste	Fermentation	Bacillus subtilis and Pseudomonas ac	surfactant	Ramírez et al., (2015
16.	Cane molasses	Fermentation and biosynthesis	Alcaligenes sp. NCIM 5085		Tripathi et al., 2019
	IN ULTICITI-FICTI 1000 WASIE	rty ur oiysus, termentauon and Biosynthesis	наютонсклучтонеттаны, наютонскосатраненые	Poity outoxybuilyrate (Prib)	Isang et al., 2019
18.	Dairy waste	Fermentation and biosynthesis	Bacillus megaterium SRKP-3	Polyh vrate (PHB)	Pandian et al. 2010, Tsang et al., 2019
19.	Food waste	Anaerobic digestion, aerobic composting	Microbial cells	rganic rer and solid	Ma et al., 2019, More et al., 2017

glucose follows two major pathways, viz. butyrate and acetate pathways that utilize one mole of glucose, yielding two and four moles of biohydrogen, respectively (Karthikeyan et al., 2018). Mixed cultures of Enterobacter and Clostridium produce hydrogen from waste, which is readily utilized by hydrogenotrophic bacteria (Li and Fang 2007). Seed biomass is heated to reduce microbial biohydrogen consumers, while suppressing lactate production, thus increasing biohydrogen production. Untreated FW abundantly contain lactic acid bacteria whereas pretreated FW is dominated by biohydrogen producers (Kiran et al., 2014).

3.1.4. Methane

Anaerobic production of methane is favored because of its renewable energy source utilization, low residual waste production and low cost. The waste produced through this process is nutrient rich that can be used as soil conditioner or fertilizer (Kiran et al., 2014). The energy content of methane is 55.5 MJ/kg. Anaerobic digestion produces methane through biodegrading and reducing organic waste. The production of methane by this process is affected by several parameters such as alkalinity, pH, organic loading rate, nutrients, reactor type, volatile fatty acids, carbon/nitrogen ratio, operation temperatures, ammonium ions, and substrate characteristics (Park et al., 2018). A decrease in pH and volatile fatty acids accumulation minimizes the production of methane gas during anaerobic digestion (Chen et al., 2008). Park et al. (2018) reported that the use of microbial electrolysis cell (MEC) with anaerobic digestion reactor enhances the methane production rate by 1.7 times as compared to the anaerobic digestion reactor alone. MEC increases the rate of degradation of volatile fatty acids, concentrated organic wastes and non-degradable organic matter thereby improving methane production. During this process, methane is formed at cathode by electrons released from exoelectrogenic bacteria when MEC pa low voltage in the reactor.

Pretreatment of FW could be an efficient strategy that aids in teins/lipids digestibility, reducing acidification rate, altering biologi and physico-chemical properties, thus avoiding proc ion an improving the recovery of methane. The pretre nt pr ss comd alka monly includes physical (grinding), thermal, ag nent. high-pressure treatment, pulse discharge of high mediated, micro-aeration and biological ment (keyan et al., kali treatm 2018). Thermal treatment followed the best pretreatpretreatment strategies for anaerob of FW. Alk 12 ment enhanced the yield of methane by 25% reas when combined with thermal treatment, the d further en d to 32% (Naran et al., 2016). FW of 54 erent fruits and etables produced , volat 180-732 mL methane olid. Furthermore fruits and veobic digester yielded 530 mL per g getables waste in a two volatile solid by utilizing 95 olatile s (Kiran et al., 2014). Zhao et al., 2017a) the e sence of varying salt (NaCl) hey obs that low salt concentration inconcentration n FW cidific nd hydrolysis processes while inhibiting creased methanoge ncentrations inhibited both methanoation. genesis and ac

3.1.5. Biobutanol

Butanol, a four carbon alcohol is considered more advanced biofuel compared to ethanol as it offers several advantages such as improved combustion efficiency, higher energy density, lower vapour pressure and property to dissolve in vegetable oils in order to reduce their viscosity (Girotto et al., 2015). Traditionally, cane molasses and starch (corn, potato, and wheat) are used for the production of biobutanol through fermentation process. Economic viability of biobutanol production via fermentation process is mainly influenced by the substrate cost, which accounts for about 50% of the cost of production (Ujor et al., 2014). Several studies have reported the efficiency of Clostridium species for the production of biobutanol using FW as a substrate (Girotto et al., 2015; Huang et al., 2015; Ujor et al., 2014). Clostridium acetobutylicum produced biobutanol through fermentation process by utilizing FW. Lactose-rich waste whey yielded 0.3 g of butanol per g of carbohydrates (Girotto et al., 2015). Similar yields of biobutanol were observed from starch-rich industrial FW such as bread liquid, batter liquid and inedible dough (Ujor et al., 2014). Clostridium beijerinckii strain P260 utilized 81 g/L of FW procured from retail store in Illinois, USA containing white bread, sweet corn, and mashed potatoes to produce biobutanol with a yield of 0.38 g/g and a high productivity of 0.46 g/L/h (Huang et al., 2015). Jesse et al., (2002) reported varying production of acetone:butanol:ethanol (ABE) by Clostridium beijerinckii strain BA101 upon action on different FW type viz. agricultural waste vields 20.3 g/L and starch-rich packing g/L whereas control starch yielded 24.7 g/L ABE. Th ference i luction could be attributed to the difference in the n of the wast terial. Since tities of 900a

employing FW showed reasonable valorization for biobutanol pr eff tion can waste management, promo c viab econo

butanol, FW t strategy for

3.2. Electric power ation

disposed through conventional Since mo generated methods like incine compost, and landfill, which were uneconor nd unsusta processes, leading to toxic gas emission (Paritosh et al., 2017). Thus, recovering ater contaminal or y or electricity employing FW is considered as eco-friendly, ecoer pollution ducing and sustainable approach. This can be n d by ana ically treating the FW in specific devices such as ac . امریک MFC) (Li et al., 2016). MFC utilizes microorganisms micr covery of electricity generated by using diverse wastes as cataly ling industrial wastewater domestic wastewater and excess sludge 2016; Cercado-Quezada et al., 2010). Briefly in MFC, microrganisms oxidize organic matter transferring electrons to anode and at cathode the oxidized compounds or oxygen get reduced microbially or by abiotic process (Cercado-Quezada et al., 2010). Organic matter rich FW serves as energy source for electricigens in MFC, hydrolysis of this organic fraction is the rate limiting step in electricity production (Li et al., 2016). Pretreatment of FW employing microwave and sonication was reported to further enhance substrate hydrolysis, which in turn increases the efficiency of electricity generation (Yusoff et al., 2013). MFC containing FW leachate at a concentration of 5000 mg/L produced maximum power 15.14 W/m³ and open circuit voltage of 1.12 V. Power output reduced with increasing substrate concentration 20,000 mg/L; this reduction was attributed to anode chamber microbial inhibition (Rikame et al., 2012). MFC performance was significantly decreased by the deposition of cations and microorganisms on fouled membranes. Jia et al., (2013) reported the microbial community structure in MFC loaded with FW collected from canteen of Harbin Institute of Technology. They have reported a maximum 18 W/m³ power density at a chemical oxygen demand (COD) of 3200 \pm 400 mg/L. The 454 pyrosequencing of the amplified 16S rRNA gene revealed the presence of different genera in anode biofilm of which fermentative Bacteroide and exoelectrogenic Geobacter were the dominant ones. Food industry wastes such as wine lees, vogurt waste and fermented apple juice in combination with two inoculums sources garden compost and anaerobic sludge was evaluated for electricity production by MFC (Cercado-Quezada et al., 2010). Of these, only yogurt waste inoculated with compost leachate exhibited the generation of stable power density of 44 mW/m². Organic fraction rich composite FW from canteen of CSIR-Indian Institute of Chemical Technology (IICT), India generated electricity upon action of anaerobic consortia as anodic biocatalyst in MFC. It was shown that OLR significantly affected the production of electricity from FW. OLR at 1.01, 1.74, and 2.61 kg COD/m³-day generated 188, 295, and 250 mV electricity, respectively (Goud et al., 2011).

4. Bio-conversion of food waste to value added products

4.1. Biosurfactants

Biosurfactants are surface active compounds of biological origin prominently produced by diverse microorganisms including bacteria, fungi, and yeast (Gaur et al., 2019a; Gaur et al., 2019b). Globally, they are estimated to generate a revenue of over 18 Billion USD with a market of 30.64 Billion USD in 2016 (Singh et al., 2019). The compound annual growth rate of biosurfactant market is estimated to be 5.6% from 2017 to 2022. This increasing demand impulse to explore cheap waste material as substrates thus reducing the production cost while obligating waste management. Thus, FW from house-holds and various food industries can be employed as substrate for biosurfactant production aiding in curtailing the production cost while diminishing the pollution. One of the prominent FW produced from house-holds and commercial kitchens is used waste oil. Kitchen waste oil (KWO) is rich in protein and moisture content, thus encourages microbial growth. Pseudomonas aeruginosa isolated from KWO preferentially utilized KWO over glucose, glycerol, molasses, and rapeseed oil as a fermentation substrate for biosurfactant production (Chen et al., 2018). Another major waste from Indian food (dairy) industry is from paneer production. With the annual production of 0.15 million tonnes of paneer generated two million tonnes of waste whey, which is often released in the environment without any pretreatment that leads to soil and water pollution (Parashar et al., 2016; Patowary et al., 2016). Pseudomonas aeruginosa SR17, a hydrocarbon-contaminated soil isolate utilized paneer whey to yield 2.7 g/L of biosurfactant (Table. 2). The production further enhanced to 4.8 g/L by additionally supplementing it with mineral salts and glucose (Patowary et al., 2016). Another strain of Pseudomonas aeruginosa ATCC 10,145 utilized soy molasses a carbon source for biosurfactant production (Rodrigues et al., 201 strain of *Pseudozyma* sp. produced biosurfactant which showed po tial application as laundry detergent additives (Saina et al 2013). S molasses is generated during soyabean processing, w com mercial value, although rich in proteins, carbol ates, lipids .7 g/ Pseudomonas aeruginosa ATCC 10,145 produ olipid biosurfactant by utilizing 120 g soy molasses s fe Pseudomonas Ramírez et al., (2015) reported that Bag subtilis aeruginosa utilized olive oil mill waste to synthesiz urfactant. These bacteria utilize remaining supplement carbon source, whereas other waste components set nutrients.

4.2. Bio-plastics

Plastics are traditio synthe zed from petrochemicals through irreversible sang et 2019). Petroleum derived proces polymers are deg by croorganisms and their high a concerns thus an alternative to persistence s seri environ bioplastics comes into play. As microic play these syn organisms e products can synthesize these bioastics are preferred over synthetic ones delineating plastics. These environmental bu while being ecofriendly. Substitution with bioplastics also offers a action in global warming concern as the energy requirement for petroleum based synthetic plastic production (77 MJ/ kg) is more as compared to bioplastics (57 MJ/kg) (Gironi and Piemonte, 2011). Land-filling of FW leads to undesirable outcomes such as groundwater contamination and greenhouse gas emissions, thus bioconversion of FW to generate bioplastics is an optimal strategy for waste disposal. The bioconversion of FW to plastics requires pretreatment to enhance biological and physico-chemical properties. The pretreatment strategies include physical, chemical, biological and enzymatic hydrolysis. Physical treatment converts FW into fermentable organic compounds employing thermal and mechanical processes, including heating, milling, ultrasound, and microwaves (Tsang et al., 2019). Fermentable sugars are formed using chemical treatment, which includes acid or alkali treatment. Biological treatment includes microorganism to utilize FW as fermentable substrate. Polyhydroxyalkanoate (PHA), hydroxybutyrate (PHB), polybutylene succinate (PBS), starch blends, polyvinyl alcohol (PVA) and polylactic acid (PLA) are the major biodegradable polymers (Prabisha et al., 2015; Pandey and Soccol, 1998; John et al., 2007; Ramachandran et al., 2007; Tripathi et al., 2019; Tsang et al., 2019).

Alcaligenes sp. NCIM 5085 through an optimized fermentation process (7.5 L) utilized cane molasses and yielded 70.89% of high molecular weight PHB with 0.312 g/L/h productivity (Tripathi et al., 2019). Another strain Halomonas campaniensis strain LS21 grew in cellulose, starch, fatty acids, fats, and proteins rich FW and produced 70% PHB at 37 °C. Bacillus megaterium SRKP-3 utilized to produce PHB (Pandian et al., 2010, Tsang et al., 201 elix et a (15) reported the production of bioplastics from free ter red swa crayfish. As per the United States Department ure, 45% the Crayfish of Ag entering in United States of App a food ma in waste only. esu¹ Guidelines by European Up the p al use of FW as suggest animal feed, but disease rol co ns made h filegal. Thus, valorizing FW by converti r it dded p ucts holds an ideal end use practice.

4.3. Organic

nal inorga artilizers contribute to increased atmo-Co anic fertilizers are considered effective sph hane emission; ative improving yield of crops along with reduction in methane al arket research, it has been estimated that emerging eı ion. As per of organi rtilizer would touch a value of over \$150–109 per m 202 W has been traditionally used as animal feed and anni r prepared by composting or vermicomposting (Du organic 2018). Agriculture waste has been utilized for mushroom cultia substrate (Philippoussis et al., 2006). It reduces environnental burden and also at the same time enhances crop productivity and changes soil bacterial community. Residue of biogas production from FW can be used as organic fertilizers or soil conditioner due to nutritional and carbon content along with macronutrients (N, K, P, Ca, Mg) and microelements (Fe, Cu, Zn and Al) capable of improving and stimulating plant growth (Zhu et al., 2015).

Organic fertilizers can be synthesized from FW by employing several processes such as anaerobic digestion, aerobic composting using microbes, chemical hydrolysis method (treating FW via alkaline or acid hydrolysis at 600-1000C), and in-situ degradation of natural organic matter. It produces fertilizers in the form of digestate / soil conditioner, compost, soluble bio-waste compost (SBC), degraded crop, minerals and as liquid organic fertilizers (Du et al., 2018). Liquid organic fertilizers (LOF) are directly delivered at root zone of individual plant in irrigation system and are therefore, advantageous in plant growth than other fertilizers. They are readily available for absorption by plant and the quantity required per field is also less, as well as the degradation process of LOF is quite easier (More et al., 2017). The efficacy of organic fertilizers can further be improved by combining methane oxidizing bacteria such as methanotrophs (Singh and Strong 2016). Organic fertilizers can support the growth of microalgae also. It can be used as an alternative nutrient medium to cultivate Chlorella vulgaris for synthesis of biodiesel (Lam and Lee, 2012).

5. Conclusions

Food is an indispensable commodity contributing to major section of organic waste generated worldwide. Improper management of FW leads to environmental hazards by releasing toxic components viz. greenhouse gases, nitrates, ammonia etc.. Presently, microbial processing of FW offers environmental protection and a sustainable way for the genesis of billion dollar revenue. It can be concluded from this review that microorganisms can be efficiently employed for biotransforming FW into complex biomolecules, bio-fertilizers, biofuels, biochemicals and electrical energy. As FW is a rich source of nutrients, its valorization is proven to be a promising approach towards FW management.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgement

The authors do not received any financial support for publication of this article.

References

- Adhikari, B., Chae, M., Bressler, D., 2018. Utilization of Slaughterhouse Waste in Value Added Applications: Recent Advances in the Development of Wood Adhesives. Polymers 10 (2), 176.
- Anal, A.K., 2017. Food Processing By-Products and their Utilization: Introduction. Food Process. By-Prod. Utiliz. 1-10.
- Ashayerizadeh, O., Dastar, B., Samadi, F., Khomeiri, M., Yamchi, A., Zerehdaran, S., 2017. Study on the chemical and microbial composition and probiotic characteristics of dominant lactic acid bacteria in fermented poultry slaughterhouse waste. Waste Manage. 65, 178–185.
- Benjamin, S., Pandey, A., 1997. Coconut cake-a potent substrate for the production of lipase by Candida rugosa in solid-state fermentation. Acta Biotechnol. 17 (3), 241-251.
- Bogar, B., Szakacs, G., Linden, J.C., Pandey, A., Tengerdy, R.P., 2003a. Optimization of phytase production by solid substrate fermentation. J. Ind. Microbiol. Biotechnol. 30 (3), 183-189.
- Bogar, B., Szakacs, G., Pandey, A., Abdulhameed, S., Linden, J.C., Tengerdy, R.P., Production of Phytase by Mucorracemosus in Solid-State Fermentation. Biot Prog. 19 (2), 312-319.
- Bolivar-Telleria, M., Turbay, C., Favarato, L., Carneiro, T., de Biasi, R.S., Fernande A.A.R., Santos, A., Fernandes, P., 2018. 2018. Second-generation bioethanol fro coconut husk, BioMed research international.
- Cercado-Quezada, B., Delia, M.L., Bergel, A., 2010. Testing vario wast for electricity production in microbial fuel cell. Bioresour 101. 1 2748-2754.
- Chang, F.C., Tsai, M.J., Ko, C.H., 2018. Agricultural wasted waste cooking oil. Environ. Sci. Pollut. Res. 25 (6)
- Chen, C., Sun, N., Li, D., Long, S., Tang, X., Xiao, ptimization and 1g. L., characterization of biosurfactant production kitchen wast ng Pseudomonas aeruginosa. Environ. Sci. Pol 5 (15), 14934
- Chen, H., Shen, H., Su, H., Chen, H., Tan, H. High-efficiency conversion tail procedure. Bioresour. of kitchen garbage to biobutanol usi enzym Technol. 245, 1110-1121.
- Chen, J., Ma, X., Yu, Z., Deng, T. X., Chen, L., Dai, N A study on catalytic TG-FTIR and Py-GC/ co-pyrolysis of kitchen was tire w e over ZSM-5 u MS. Bioresour, Technol.
- nibition anaerobic digestion process: a Chen, Y., Cheng, J.J., Cream review. Bioresour. Technol 044-4064 waste (as a renewable source of energy: icip

s in Ò

Pinto, C

do 1

tech

Cheng, H., Hu, Y., 20 Current and f Cubas ALV

auction Biodiese charge Dai, Y., Sun,

Zheng, W.,

contaminants:

Dede, O.H., Ozdemir,

a food industry waste using corona disfatty acids ge. 47, 149–154. I., Li, J., Yang, S., Sun, Y., Zhang, K., Xu, J., lizations of agricultural waste as adsorbent for the removal of Chemosphere 211, 235-253.

fuel f

. Technol. 101 (11), 3816-3824.

oecke, E.H.S., Dutra, A.R.A., 2016.

- Development of nutrient-rich growing media with ha-
- zelnut husk and mur 1 sewage sludge. Environ. Technol. 39 (17), 2223-2230. Dessie, W., Zhang, W., Xin, F., Dong, W., Zhang, M., Ma, J., Jiang, M., 2018. Succinic acid production from fruit and vegetable wastes hydrolyzed by on-site enzyme mixtures
- through solid state fermentation. Bioresour. Technol. 247, 1177-1180. Dias, T., Fragoso, R., Duarte, E., 2014. Anaerobic co-digestion of dairy cattle manure and pear waste. Bioresour. Technol. 164, 420-423.
- Díaz, A.I., Laca, A., Laca, A., Díaz, M., 2017. Treatment of supermarket vegetable wastes to be used as alternative substrates in bioprocesses. Waste Manage. 67, 59-66.
- Diwan, B., Parkhey, P., Gupta, P., 2018. From agro-industrial wastes to single cell oils: a step towards prospective biorefinery. Folia Microbiol. 63 (5), 547-568.
- Du, C., Abdullah, J.J., Greetham, D., Fu, D., Yu, M., Ren, L., Li, S., Lu, D., 2018. Valorization of food waste into biofertiliser and its field application. J. Cleaner Prod. 187 273-284
- Felix, M., Romero, A., Cordobes, F., Guerrero, A., 2015. Development of crayfish biobased plastic materials processed by small-scale injection moulding. J. Sci. Food Agric. 95 (4), 679-687.

Fewtrell, L., 2004. Drinking-water nitrate, methemoglobinemia, and global burden of disease: a discussion. Environ. Health Perspect. 112 (14), 1371-1374.

- Galanakis, C.M., 2017. Handbook of Grape Processing by-Products: Sustainable. Solutions. Academic Press
- Gaur, V.K., Bajaj, A., Regar, R.K., Kamthan, M., Jha, R.R., Srivastava, J.K., Manickam, N., 2019a. Rhamnolipid from a Lysinibacillus sphaericus strain IITR51 and its potential application for dissolution of hydrophobic pesticides. Bioresour. Technol. 272, 19-25.
- Gaur, V.K., Regar, R.K., Dhiman, N., Gautam, K., Srivastava, J.K., Patnaik, S., Kamthan, M., Manickam, N., 2019b. Biosynthesis and characterization of sophorolipid biosurfactant by Candida spp.: application as food emulsifier and antibacterial agent. Bioresour. Technol. 285, 121314.
- Gironi, F., Piemonte, V., 2011. Bioplastics and petroleum-based plastics: strengths and weaknesses. Energy Sources Part A 33 (21), 1949-1959.
- Girotto, F., Alibardi, L., Cossu, R., 2015. Food waste generation and industrial uses: a review. Waste Manage. 45, 32-41.
- Goud, R.K., Babu, P.S., Mohan, S.V., 2011. Cantee ood waste as potential anodic fuel for bioelectricity generat single cha microbial fuel cell (MFC): bio-electrochemical evaluation loading conincreasing su dition. Int. J. Hydrogen Energy 36 (10), 62
- of a genetically Heater, B.S., Chan, W.S., Lee, M.M., Cha (.. 201 ted evol encoded immobilized lipase for t cient prod from waste (1), 16cooking oil. Biotechnol. Biofu
- 1., Lova Henkel, M., Müller, M.M., Kügl .B., Cont ., Syldatk, C., Hausmann, R., 2012. Rhai arfactants from renewable resources: concepts for next-gen producti rocess Biochem. 47 (8), 1207–1219.
- Heredia-Guerrero, J ngolani, R., Bayer, I.S., eredia, A., D J.J., 2017. C Athanassiou agro-waste as a raw material for the . Exp. Bot. 68), 5401–5410. production
- Hong, Y.S., Yoon H.H., 20 anol production from food residues. Biomass Bioenergy 35 (7) 3271-3275.
- , V., Oureshi, I Huan . Butanol production from food waste: a novel ss for producing sustaina energy and reducing environmental pollution. technol. Biofuels 8 (1), 147.
 - n. M.H., El-Ha J.F., Shehata, H.A., Hegazy, M.A., Hefni, H.H., 2013.
 - aration of so co-friendly corrosion inhibitors having antibacterial activity ea food w J. Surfactants Deterg. 16 (2), 233-242.
 - muak, C., Putmai, N., Pavasant, P., 2018. Scaling-up bio-hydrog n from food waste: Feasibilities and challenges. Int. J. Hydrogen Energy 43 (2), 634-648.
 - V., Ezeji, T.C., Qureshi, N., Blaschek, H.P., 2002. Production of butanol from sed waste packing peanuts and agricultural waste. J. Ind. Microbiol. Biotechnol. 29 (3), 117-123.
- Ji, C., Kong, C.X., Mei, Z.L., Li, J., 2017. A review of the anaerobic digestion of fruit and vegetable waste. Appl. Biochem. Biotechnol. 183 (3), 906-922.
- Jia, J., Tang, Y., Liu, B., Wu, D., Ren, N., Xing, D., 2013. Electricity generation from food wastes and microbial community structure in microbial fuel cells. Bioresour, Technol. 144 94-99
- John, R.P., Nampoothiri, K.M., Pandey, A., 2007. Fermentative production of lactic acid from biomass: an overview on process developments and future perspectives. Appl. Microbiol. Biotechnol. 74 (3), 524-534.
- Karmee, S.K., 2016. Liquid biofuels from food waste: current trends, prospect and limitation. Renew. Sustain. Energy Rev. 53, 945-953.
- Karmee, S.K., Linardi, D., Lee, J., Lin, C.S.K., 2015. Conversion of lipid from food waste to biodiesel. Waste Manage. 41, 169-173.
- Karthikeyan, O.P., Trably, E., Mehariya, S., Bernet, N., Wong, J.W., Carrere, H., 2018. Pretreatment of food waste for methane and hydrogen recovery: a review. Bioresour. Technol. 249, 1025-1039.
- Kiran, E.U., Trzcinski, A.P., Ng, W.J., Liu, Y., 2014. Bioconversion of food waste to energy: a review. Fuel 134, 389-399.
- Kondamudi, N., Mohapatra, S.K., Misra, M., 2008. Spent coffee grounds as a versatile source of green energy. J. Agric. Food. Chem. 56 (24), 11757-11760.
- Kumar, A., Kumar, D., George, N., Sharma, P., Gupta, N., 2018. A process for complete biodegradation of shrimp waste by a novel marine isolate Paenibacillus sp. AD with simultaneous production of chitinase and chitin oligosaccharides. Int. J. Biol. Macromol. 109, 263-272.
- Kumar, S., Sangwan, P., Dhankhar, R.M.V., Bidra, S., 2013. Utilization of rice husk and their ash: A review. Res. J. Chem. Env. Sci 1 (5), 126-129.
- Kushwaha, J.P., Srivastava, V.C., Mall, I.D., 2011. An overview of various technologies for the treatment of dairy wastewaters. Crit. Rev. Food Sci. Nutr. 51 (5), 442-452.
- Lam, M.K., Lee, K.T., 2012. Potential of using organic fertilizer to cultivate Chlorella vulgaris for biodiesel production. Appl. Energy 94, 303-308.
- Lappa, I.K., Papadaki, A., Kachrimanidou, V., Terpou, A., Koulougliotis, D., Eriotou, E., Kopsahelis, N., 2019. Cheese Whey Processing: Integrated Biorefinery Concepts and Emerging Food Applications. Foods 8 (8), 347.
- Li, C., Fang, H.H., 2007. Fermentative hydrogen production from wastewater and solid wastes by mixed cultures. Critical Reviews in Environmental Science and Technology 37 (1), 1-39
- Li, H., Tian, Y., Zuo, W., Zhang, J., Pan, X., Li, L., Su, X., 2016. Electricity generation from food wastes and characteristics of organic matters in microbial fuel cell. Bioresour. Technol. 205, 104-110.
- Li, P., Zeng, Y., Xie, Y., Li, X., Kang, Y., Wang, Y., Xie, T., Zhang, Y., 2017. Effect of pretreatment on the enzymatic hydrolysis of kitchen waste for xanthan production. Bioresour. Technol. 223, 84–90.
- Liu, W., Dong, Z., Sun, D., Chen, Y., Wang, S., Zhu, J., Liu, C., 2019. Bioconversion of

H

Jarui

P. Sharma, et al.

kitchen wastes into bioflocculant and its pilot-scale application in treating iron mineral processing wastewater. Bioresour. Technol. 288, 121505.

- Ma, H., Wang, Q., Qian, D., Gong, L., Zhang, W., 2009. The utilization of acid-tolerant bacteria on ethanol production from kitchen garbage. Renewable Energy 34 (6), 1466-1470.
- Ma, J., Pan, J., Qiu, L., Wang, Q., Zhang, Z., 2019. Biochar triggering multipath methanogenesis and subdued propionic acid accumulation during semi-continuous anaerobic digestion. Bioresour. Technol. 122026.
- Madurwar, M.V., Ralegaonkar, R.V., Mandavgane, S.A., 2013. Application of agro-waste for sustainable construction materials: A review. Constr. Build. Mater. 38, 872-878. Mahboubi, A., Ferreira, J.A., Taherzadeh, M.J., Lennartsson, P.R., 2017. Value-added
- products from dairy waste using edible fungi. Waste management (New York, NY) 59, 518.
- Marques, R.V., Paz, M.F.D., Duval, E.H., Corrêa, L.B., Corrêa, É.K., 2016. Staphylococcus xylosus fermentation of pork fatty waste: raw material for biodiesel production. Braz. J. Microbiol. 47 (3), 675-679.
- Matharu, A.S., de Melo, E.M., Houghton, J.A., 2016. Opportunity for high value-added chemicals from food supply chain wastes. Bioresour. Technol. 215, 123-130.
- More, A., Srinivasan, A., Liao, P.H., Lo, K.V., 2017. Microwave enhanced oxidation treatment of organic fertilizers. J. Sci. Food Agric. 97 (10), 3233-3239.
- Murthy, P.S., Naidu, M.M., 2012. Sustainable management of coffee industry by-products and value addition-A review. Resour. Conserv. Recycl. 66, 45-58.
- Naran, E., Toor, U.A., Kim, D.J., 2016. Effect of pretreatment and anaerobic co-digestion of food waste and waste activated sludge on stabilization and methane production. Int. Biodeterior. Biodegrad. 113, 17-21.
- Nikolaidis, P., Poullikkas, A., 2017. A comparative overview of hydrogen production processes. Renew. Sustain. Energy Rev. 67, 597-611.
- Ning, Z., Zhang, H., Li, W., Zhang, R., Liu, G., Chen, C., 2018. Anaerobic digestion of lipid-rich swine slaughterhouse waste: Methane production performance, long-chain fatty acids profile and predominant microorganisms. Bioresour. Technol. 269, 426-433.
- Nishimura, H., Tan, L., Kira, N., Tomiyama, S., Yamada, K., Sun, Z.Y., Tang, Y.Q., Morimura, S., Kida, K., 2017. Production of ethanol from a mixture of waste paper and kitchen waste via a process of successive liquefaction, presaccharification, and simultaneous saccharification and fermentation. Waste Manage. 67, 86-94.
- Okino-Delgado, C.H., Do Prado, D.Z., Facanali, R., Marques, M.M.O., Nascimento, A.S., da Costa Fernandes, C.J., Zambuzzi, W.F., Fleuri, L.F., 2017. Bioremediation of cooking oil waste using lipases from wastes. PLoS ONE 12 (10), e0186246.
- Panda, S.K., Mishra, S.S., Kavitesi, F., Ray, R.C., 2016, Microbial-processing of fruit and vegetable wastes for production of vital enzymes and organic acids: Biotech and scopes, Environ, Res. 146, 161-172.
- Pandey, A., Soccol, C.R., 1998. Bioconversion of biomass: a case study of ligno-cell bioconversions in solid state fermentation. Brazilian Archives of Biology and Technology 41 (4), 379-390.
- Pandey, A., Soccol, C.R., 2000. Economic utilization of crop residue ddition futuristic approach. J. Sci. Ind. Res. 59, 12-22. Pandian, S.R., Deepak, V., Kalishwaralal, K., Rameshkumar, M raj, M. unathar
- S., 2010. Optimization and fed-batch production of PHP ing dair te and sea water as nutrient sources by Bacillus megaterium SRK 705-711.
- Papageorgiou, M., Skendi, A., 2018. Introduction t 1 process by-products. In Sustainable Recovery and Reutilization of (rocessing By-Woodhead Publishing, pp. 1-25.
- Papanikolaou, S., Dimou, A., Fakas, S., Dia Philippoussis, ... Galiotouopo technological Panayotou, M., Aggelis, G., 2011. Bi on of waste cooking olive oil into lipid-rich biomass using dlus and Penicil ins. J. Appl. Microbiol. 110 (5), 1138–1150.
- e, M., B r, D.C., 2016. Parashar, A., Jin, Y., Mason, B acorporation of whey ethano permeate, a dairy efflue entation to provide a zero waste solution 3), 1859for the dairy industry. J.
- ation of t Parate, V.R., Talib, M.I., 2015. C (Cajanuscajan) husk carbon and its kinetics and study oving 1) ions, IOSR J. Environ, Sci. Toxicol. Foo 7-41.
- tha, S.F N., Chawade, A., Vivekanand, V., 2017. Paritosh, K., I adav, M., o energ tainable approaches for food waste man-Food v erview of s ed research international. agemen ıtri Park, J., Lee, B . Bioelectrochemical enhancement of methane

production fi

- ly concentrated food waste in a combined anaerobic digester and microbial e s cell. Bioresour. Technol. 247, 226-233.
- Patowary, R., Patowary, ita, M.C., Deka, S., 2016. Utilization of paneer whey waste for cost-effective prod ion of rhamnolipid biosurfactant. Appl. Biochem. Biotechnol. 180 (3), 383-399.
- Philippoussis, A.N., 2009. Production of mushrooms using agro-industrial residues as substrates. In Biotechnology for agro-industrial residues utilisation. Springer, Dordrecht, pp. 163-196.
- Pleissner, D., Lam, W.C., Sun, Z., Lin, C.S.K., 2013. Food waste as nutrient source in heterotrophic microalgae cultivation. Bioresour. Technol. 137, 139-146.
- Prabisha, T.P., Sindhu, R., Binod, P., Sankar, V., Raghu, K.G., Pandey, A., 2015. Production and characterization of PHB from a novel isolate Comamonas sp. from a dairy effluent sample and its application in cell culture. Biochem. Eng. J. 101, 150-159.
- Prameela, K., Venkatesh, K., Immandi, S.B., Kasturi, A.P.K., Krishna, C.R., Mohan, C.M., 2017. Next generation nutraceutical from shrimp waste: The convergence of applications with extraction methods. Food Chem. 237, 121-132.
- Ramachandran, S., Singh, S.K., Larroche, C., Soccol, C.R., Pandey, A., 2007. Oil cakes and their biotechnological applications-A review. Bioresour. Technol. 98 (10),

2000-2009.

Si

Sin

Singh, S

- Ramírez, I.M., Tsaousi, K., Rudden, M., Marchant, R., Alameda, E.J., Román, M.G., Banat, I.M., 2015. Rhamnolipid and surfactin production from olive oil mill waste as sole carbon source. Bioresour. Technol. 198, 231-236.
- Rikame, S.S., Mungray, A.A., Mungray, A.K., 2012. Electricity generation from acidogenic food waste leachate using dual chamber mediator less microbial fuel cell. Int. Biodeterior. Biodegrad. 75, 131-137.
- Rodrigues, M.S., Moreira, F.S., Cardoso, V.L., de Resende, M.M., 2017. Soy molasses as a fermentation substrate for the production of biosurfactant using Pseudomonas aeruginosa ATCC 10145. Environ. Sci. Pollut. Res. 24 (22), 18699-18709.
- Rohm, H., Brennan, C., Turner, C., Günther, E., Campbell, G., Hernando, I., Struck, S., Kontogiorgos, V., 2015. Adding value to fruit processing waste: innovative ways to incorporate fibers from berry pomace in baked and extruded cereal-based foods-SUSFOOD project. Foods 4 (4), 690-697.
- Sabu, A., Sarita, S., Pandey, A., Bogar, B., Szakacs, G., Soccol, C.R., 2002. Solid-state fermentation for production of phytase by Rhizop Appl. Biochem. Biotechnol. 102 (1–6), 251–260.
- Sadh, P.K., Duhan, S., Duhan, J.S., 2018. Agroal wastes ir utilization using solid state fermentation: a review. I trces and Biop sing 5 (1), 1. Sajna, K.V., Sukumaran, R.K., Jayamurthy, H., F .K., Kanjilal, Prasad, R.B.,
- Pandey, A., 2013. Studies on biosurf ts from ozvma sp 8165 and their potential application as laundry ent additiv J. 78, 85–92. acid fer fuse using thermo-Sakai, K., Ezaki, Y., 2006. Open I
- entatio philic Bacillus coagulans ar rescent tu hvbri n analysis of microflora. J. Biosci. Bioeng. 10 157-
- Schieber, A., 2017. Side s a source of valuable comd process pounds: Selected e nce and technology 8, 97–112. ew of fo es. Anni ry, L., Pande Selvakumar, P., Ash synthesis of glucoamylase from Aspergillus nig g tea waste as the basis of a solid -state fermen substrate. ol. 65 (1-2). 55. 011
- 16. Vegetable oil deodorizer distillate: a rich source of Sherazi, S.T.H., ahesar, al bioactive cor ts. J. Oleo Sci. ess16125. the nati
- Singh ng, P.J., 2016. I ally derived fertilizer: a multifaceted bio-tool in me mitigation. Ecotoxico. nviron. Saf. 124, 267–276.
 - N.B., Singh, R., Imam, M.M., 2014. Waste water management in dairy industry: llution abaten and preventive attitudes. International Journal of Science, ennment and te logy 3 (2), 672-683. Patil Y F
 - V., 2019. Biosurfactant production: emerging trends and pro-Appl. Microbiol. 126 (1), 2-13.
 - .U., Pandey, A., 2006. Metabolic engineering approaches for lactic acid production. Process Biochem. 41 (5), 991-1000.
 - iao, Y., Liu, R., Wu, T., Zhang, M., 2019. Steam explosion modification on tea enhance bioactive compounds' extractability and antioxidant capacity of ts. J. Food Eng. 261, 51–59. extra
- Tan, L., Sun, Z., Zhang, W., Tang, Y., Morimura, S., Kida, K., 2014. Production of bio-fuel ethanol from distilled grain waste eluted from Chinese spirit making process Bioprocess Biosyst. Eng. 37 (10), 2031–2038.
- Thomsen, A.B., Medina, C., Ahring, B.K., 2003. Biotechnology in ethanol production. In Risø energy report 2. New and emerging bioenergy technologies.
- Tripathi, A.D., Raj Joshi, T., Kumar Srivastava, S., Darani, K.K., Khade, S., Srivastava, J., 2019. Effect of nutritional supplements on bio-plastics (PHB) production utilizing sugar refinery waste with potential application in food packaging. Prep. Biochem Biotech, 1–11.
- Tsang, Y.F., Kumar, V., Samadar, P., Yang, Y., Lee, J., Ok, Y.S., Song, H., Kim, K.H., Kwon, E.E., Jeon, Y.J., 2019. Production of bioplastic through food waste valorization. Environ, Int. 127, 625-644.
- Ujor, V., Bharathidasan, A.K., Cornish, K., Ezeji, T.C., 2014. Feasibility of producing butanol from industrial starchy food wastes. Appl. Energy 136, 590-598.
- Valcarcel, J., Novoa-Carballal, R., Perez-Martín, R.I., Reis, R.L., Vázquez, J.A., 2017 Glycosaminoglycans from marine sources as therapeutic agents. Biotechnol. Adv. 35 (6), 711–725.
- Verni, M., Rizzello, C.G., Coda, R., 2019,. Fermentation biotechnology applied to cereal industry by-products: nutritional and functional insights. Front. Nutr. 6.
- Wu, Y., Ma, H., Zheng, M., Wang, K., 2015. Lactic acid production from acidogenic fermentation of fruit and vegetable wastes. Bioresour. Technol. 191, 53-58.
- Yaakob, M.A., Mohamed, R.M.S.R., Al-Gheethi, A., Tiey, A., Kassim, A.H.M., 2019. Optimising of Scenedesmus sp. biomass production in chicken slaughterhouse wastewater using response surface methodology and potential utilisation as fish feeds. Environ. Sci. Pollut. Res. 26 (12), 12089-12108.
- Yan, N., Chen, X., 2015. Sustainability: Don't waste seafood waste. Nature News 524 (7564), 155.
- Yang, X., Lee, S.J., Yoo, H.Y., Choi, H.S., Park, C., Kim, S.W., 2014. Biorefinery of instant noodle waste to biofuels. Bioresour. Technol. 159, 17-23.
- Yusoff, M.Z.M., Hu, A., Feng, C., Maeda, T., Shirai, Y., Hassan, M.A., Yu, C.P., 2013. Influence of pretreated activated sludge for electricity generation in microbial fuel cell application. Bioresour. Technol. 145, 90-96.
- Zhao, J., Liu, Y., Wang, D., Chen, F., Li, X., Zeng, G., Yang, Q., 2017a. Potential impact of salinity on methane production from food waste anaerobic digestion. Waste Manage. 67, 308–314.
- Zhao, K., Xu, R., Zhang, Y., Tang, H., Zhou, C., Cao, A., Zhao, G., Guo, H., 2017b. Development of a novel compound microbial agent for degradation of kitchen waste. Braz. J. Microbiol. 48 (3), 442-450.
- Zhu, N.M., Luo, T., Guo, X.J., Zhang, H., Deng, Y., 2015. Nutrition potential of biogas residues as organic fertilizer regarding the speciation and leachability of inorganic metal elements. Environ. Technol. 36 (8), 992-1000.